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2 Phycoremediation of industrial wastewater: review of algae consortia

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6 Abstract

7 Industrialisation, population growth, and concomitant demand for fresh water have immensely impacted water quality and
8 scarcity. In particular, wastewater generated from industries generally produces high amounts of nutrients, heavy metals, and
9 chemicals that degrade the environment. Several algal species have been studied and utilized for their role in the treatment
10 of various types of wastewater. Conventional wastewater treatment options are often expensive and energy-demanding, and
11 generally ineffective at completely removing contaminants. Conversely, phycoremediation technology is an emerging green
12 approach used to remove various types of pollutants from the environment while producing valuable compounds. Compared
13 to conventional methods, phycoremediation presents as an eco-friendly and economically attractive alternative. This paper
14 serves as a review of an algal-based treatment technology in wastewater remediation for industry, describing the most com-
15 mon microalgal consortia used for this purpose. Phycoremediation challenges and strategies to urgently accelerate steps
16 towards achieving a clean and safe environment are presented, while examples of applications in industries are also provided.

17 **Keywords** Bioremediation · Industrial effluents · Metal pollution · Microalgae · Wastewater treatment

18 Introduction

19 Uncontrolled population growth, urban migration, and
20 industrialisation are among the most urgent 21st-century
21 challenges globally, exacerbated in Africa where sanitation
22 services and infrastructure are already under severe
23 pressure (UN Habitat 2008; Meso et al. 2016; Brikké and
24 Vairavamoorthy 2016; Nansubunga et al. 2016; Saghir and
25 Santoro 2018). The availability of fresh water is scarce and
26 environmental problems are more severe due to rapid indus-
27 trialisation and urbanization. The global water demand of
28 various sectors (particularly agriculture, industries, and
29 municipalities) is expected to increase by as much as 30%

by 2050 (Boretti and Rosa 2019), while the annual global
waste production is expected to increase from 2.01 to 3.40
billion tonnes (Koul et al. 2022). This will further impact
people and ecosystem services.

Conventional techniques of wastewater treatment and
remediation treatment processes involve several steps includ-
ing preliminary, primary, secondary, tertiary, and disinfection
processes; and these processes are often ineffective at
removing all contaminants present in wastewater. To negate
the harmful impact of various pollutants, alternative treat-
ment options are required. Phycoremediation using either
macroalgae (multicellular and macroscopic) or microal-
gae (unicellular and microscopic) has become a promising
method for biological purification and treatment of waste-
water (Olgui'n 2003; Gani et al. 2015). Phycoremediation
offers various advantages over other traditional methods
(e.g., low cost, carbon sequestration). Microalgae are of
increasing interest in various sectors, including renewable
energy, food, environmental management, water purifica-
tion, and the production of biofertilizers, and can provide
an alternative low-cost and low-energy solution to nutrient
recovery from wastewater streams in developing countries.

Biological treatment of wastewater, more specifically
plant-based green technologies, can assist in support-
ing the United Nations Sustainable Development Goals

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(SDGs) and contribute significantly to the circular economy (Fig. 1). Phycoremediation is aligned, in particular, to the objectives of SDG-3 “Good health and well-being”, SDG-6 “Clean water and sanitation”, SDG-7 “Affordable and clean energy”, and SDG-13 “Climate action”, collectively directed to enhance social, economic, and environmental aspects.

Internationally phycoremediation approaches have already been successfully applied to several industrial wastewater types including paper and pulp (Sasi et al. 2020), distillery (Khrisnamoorthy et al. 2019), and sugar processing (Sailaja and Meti 2014; Zewdie and Ali 2020) effluents. Similarly, phycoremediation has reportedly successfully decreased or eliminated the heavy metal content of wastewater, and reduced antibiotic resistance and other emerging contaminants (Sutherland and Ralph 2019; Leng et al. 2020; Singh et al. 2021; Jha et al. 2023; Atoku et al. 2021; Hejna et al. 2022; Ricky and Shanthakumar 2022). In South Africa specifically, phycoremediation has been applied in municipal and domestic wastewater treatment (Oberholster et al. 2017, 2021; Steyn and Oberholster 2021). Given the current water quality challenges and expected future population growth, phycoremediation presents as a sustainable wastewater management option in support of closing the loop in the circular economy. Figure 2 provides a timeline for phycoremediation globally. This review, therefore, documents the effectiveness of phycoremediation as an alternative technology to treat and remove metals and other toxic pollutants from industrial wastewater, as well as the possible opportunities and challenges facing the phycoremediation of industrial wastewater.

Mechanism of phycoremediation of microalgae

While phycoremediation is not a new concept (John 2003; Abdel-Raouf et al. 2012), it has recently gained significant attention for treating domestic, industrial, agro-industrial, and livestock wastewater (Abdelfattah et al. 2022). Bioremediation using microalgae is favoured above macroalgae due to its simple organization, faster growth rate, and high cell surface-to-volume ratio, as well as being capable of doubling their biomass in less than one day (Demirbas 2011; Lam et al. 2012; Kumar et al. 2018). As a further advantage, Leong and Chang (2020) described the high photosynthetic performance of algae and its ability to adapt to extreme environmental conditions. Bansal et al.



Fig. 1 Phycoremediation in the circular economy in support of the SDGs (adapted from Olabi et al. 2023)

(2018) reported that phycoremediation technology generally improved several biochemical parameters of water including pH, colour and odour removal, sludge reduction, biological oxygen demand (BOD) and chemical oxygen demand (COD) reduction, and nutrient removal including ammonia. The reduction of operational costs by up to 90% and better management of R/O rejects were also reported (Bansal et al. 2018). A variety of microalgal species, including *Scenedesmus*, *Chlorella*, *Botryococcus*, *Phormidium*, *Limnospira* (formerly *Arthrospira*, *Spirulina*), and *Chlamydomonas*, have been associated with successful removal of nutrients, heavy metals, emerging pollutants and pathogens from wastewater (Hoh et al. 2016; López-Sánchez et al. 2022; Ahmad et al. 2021). Studies have shown that phycoremediation can reduce eutrophication by 57.4%, ozone by 22.6%, and global warming by 2.7% (Li et al. 2020).

While the exact mechanism of phycoremediation is largely unknown, it is believed that microalgae bind pollutants using cell wall receptor molecules such as lipids, proteins, and polysaccharides (Priyadarshani et al. 2011; Balzano et al. 2020) by creating a negative charge on the cell surface (Anastopoulos and Kyzas 2015). Microalgae-facilitated bioremediation involves mechanisms (Fig. 3) such as biosorption, bioaccumulation, biodegradation, biotransformation, phycovolatilisation, and complexation (Shackira et al. 2022; Abdelfattah et al. 2022). Biosorption is a passive

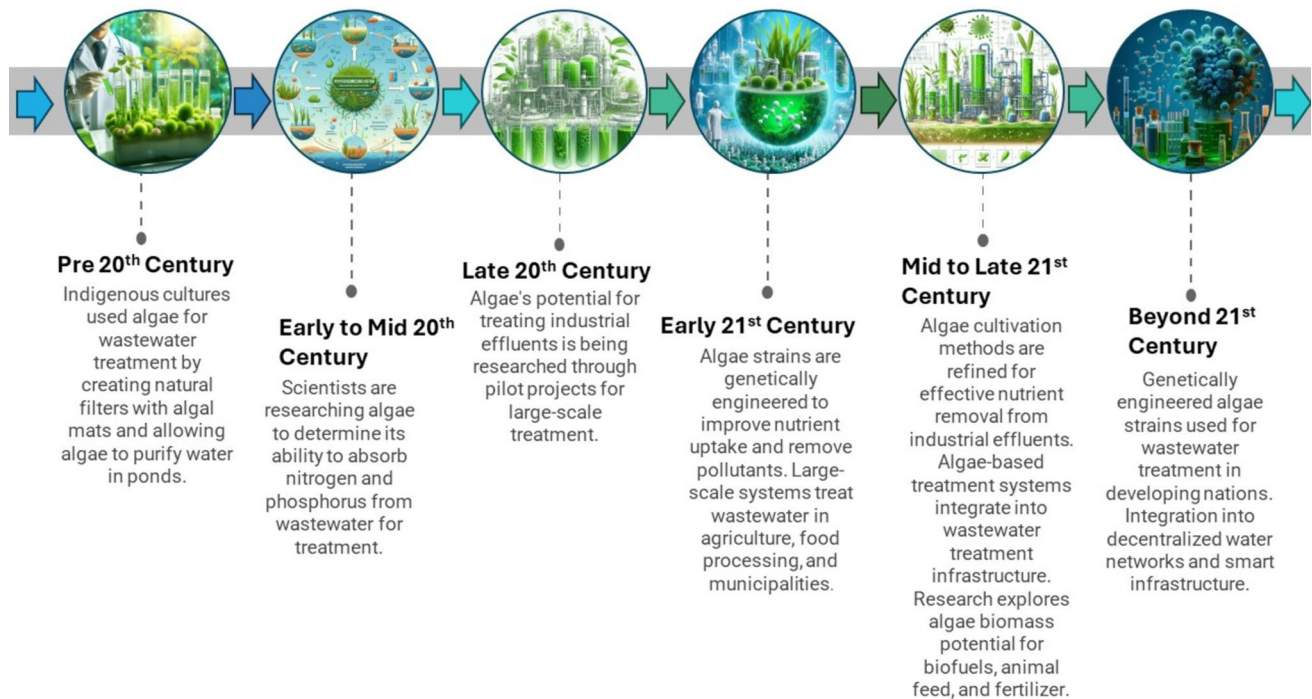


Fig. 2 Timeline of algae-based treatment of wastewater

128 biological physicochemical process whereby environmental
 129 pollutants (e.g., metal ions and dyes) can be absorbed or
 130 adsorbed (Brazesh et al. 2021). In contrast, bioaccumulation
 131 is an active treatment process that uses energy and involves
 132 the absorption of contaminants by living cells, which then
 133 accumulate or metabolize the contaminants. Furthermore,
 134 Mustafa et al. (2021) described bioaccumulation as an essen-
 135 tial pathway for the removal of organic and inorganic pol-
 136 lutants from industrial wastewater. Biodegradation is one of
 137 the most effective processes to remove contaminants from
 138 effluents, and involves the breakdown of target pollutants
 139 either through 1) biotransformation which entails a series
 140 of enzymatic reactions to produce different metabolites,
 141 or 2) through complete mineralization of the parent com-
 142 pound to CO₂ and H₂O (Xiong et al. 2017). Through the
 143 processes outlined above, microalgae are able to remove
 144 nutrients, heavy metals, dyes, and xenobiotics. In addition
 145 to bioremediation capabilities, phycoremediation presents
 146 circular economy opportunities; microalgae biomass can
 147 be harvested and transformed into economically valuable
 148 products such as biofertilizers, biofuel, biogas, and health
 149 supplements.

150 Several processes inform microalgae's potential to remove
 151 hazardous compounds from wastewater. For instance, micro-
 152 algae employ metabolic processes for pollutant removal

153 from wastewater. (Fig. 4). Some of the key metabolic pro-
 154 cesses include photosynthesis (to convert carbon dioxide
 155 (CO₂) to oxygen (O₂), respiration (taking O₂ and releasing
 156 CO₂), nutrient uptake (microalgae require nutrients such as
 157 N and P, and assimilate these nutrients into their biomass,
 158 thereby reducing nutrient concentrations in the water), bio-
 159 degradation (metabolizing organic pollutants into simpler,
 160 less toxic substances), and extracellular enzyme produc-
 161 tion (that facilitate the breakdown of organic pollutants in
 162 wastewater).

163 Selection of algae species for treating 164 industrial wastewater

165 Algae (macro- and micro-algae) are a diverse group of
 166 aquatic plants, with cellulose-based cell walls, lacking typi-
 167 cal roots, and known to have high carbon mitigating and
 168 photosynthetic efficiencies (Priyadarshini et al. 2021). The
 169 selection of suitable microalgal species (or consortia thereof)
 170 is important depending upon the characteristics of wastewa-
 171 ter for efficient wastewater treatment. As mentioned, a wide
 172 variety of algae are already applied to industrial wastewater
 173 treatment (Batista et al. 2015; Empanan et al. 2019). Table 1

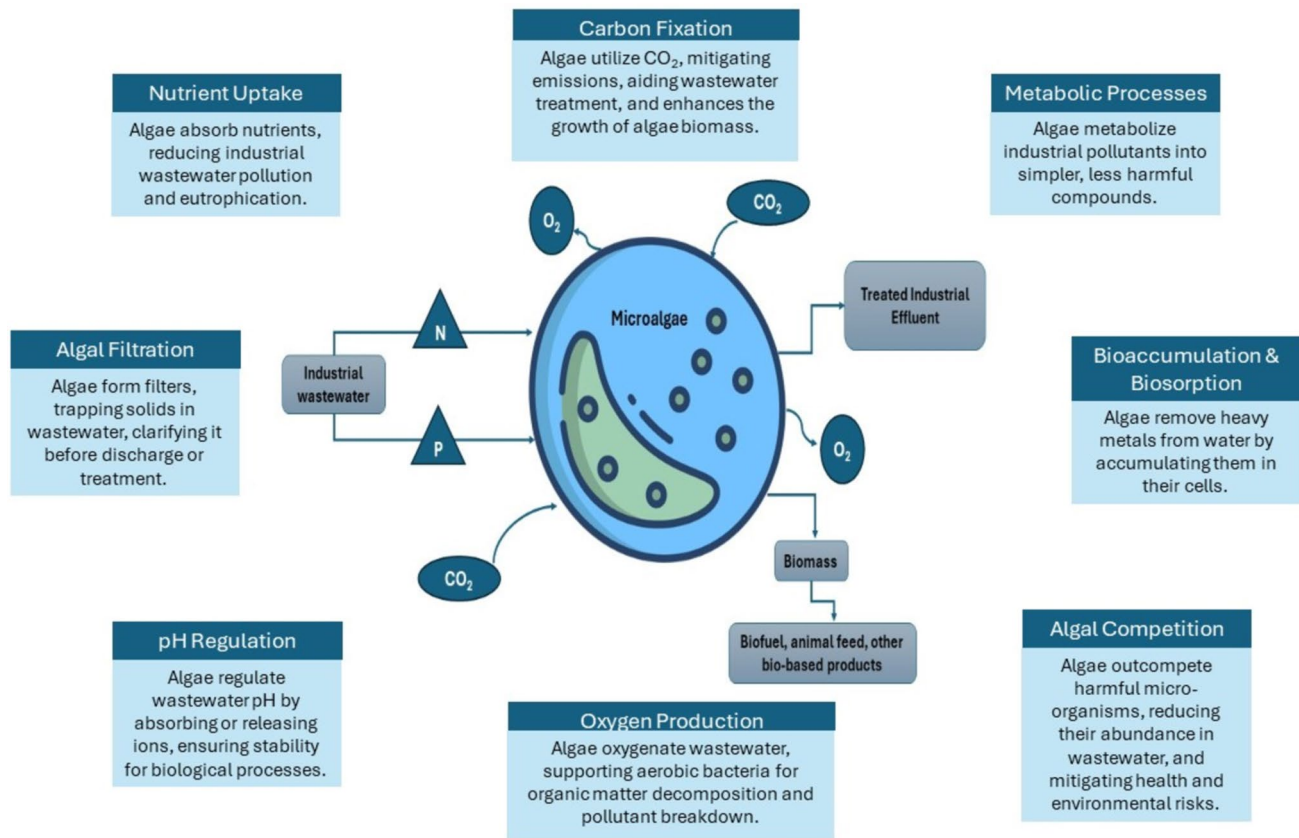


Fig. 3 Mechanisms used by algae cells to treat polluted water

174 provides a summary of laboratory-scale studies in the liter-
 175 ature describing the various microalgal species used to
 176 effectively treat industrial wastewater.

177 Microalgae have the ability to remove heavy metals in
 178 wastewater due to the presence of proteins, lipids, and poly-
 179 saccharides with heavy metal binding functional groups on
 180 their cell wall surface (Priatni et al. 2018). Reports on heavy
 181 metal affinity are numerous. Filamentous blue-green algae
 182 and freshwater filamentous diatoms have the potential to
 183 bind heavy metals effectively, while brown algae (e.g., *Asco-*
 184 *phyllum sp.*, *Sargassum sp.*) has a high affinity for divalent
 185 cations (Cd^{2+} , Zn^{2+} , Ca^{2+} , Pb^{2+} , Cu^{2+}). On the other hand,
 186 lead (Pb) is sorbed maximally by the majority of algal spe-
 187 cies when compared to other metals.

188 For organic substances, certain microalgae strains are
 189 known to successfully utilize the organic matter in efflu-
 190 ents, reducing COD, BOD, and total organic carbon levels
 191 (TOC). For example, several studies reported that *Scened-*
 192 *esmus sp.* cultivation in pulp and paper mill waste efficiently
 193 removed > 75% COD, > 82% BOD, ~ 93% NH_4 , > 65%

194 NO_3^- , ~97% PO_4^{3-} and $\text{PO}_4\text{-P}$ -71.29% (Table 1) (Usha et al. 194
 195 2016; Pham et al. 2020), while *Planktochlorella Nurekis* 195
 196 and *Clamydomonas Reinhardtii* removed 92.8% COD, 95.6% N, 196
 197 100% P, and 91.9% COD, 85.6% N, 87.5% P; respectively 197
 198 (Sasi et al. 2020). Lower removal efficiencies were reported 198
 199 for *Oocystis sp.* (71.1% COD, 70% P; Riano et al. (2011)), *C.* 199
 200 *zofingiensis* (82.70% TN, 68.96% COD; Zhu et al. (2013)), 200
 201 *Pithopora sp.* (87.23% COD, 87.75% BOD; Murugesan and 201
 202 Dhamotharan (2009)) and *Botryococcus sp.* (70.68% COD, 202
 203 61.11% BOD, 76.66% TOC, 72.49% TN, 78.23% P, 86.62% 203
 204 N; Gani et al. (2015)). 204

Phycoremediation treatment techniques 205

206 Conventional treatment of industrial wastewater includes 206
 207 various chemical and physical treatment techniques. These 207
 208 techniques typically include chlorination, coagulation/floc- 208
 209 culation, ultraviolet light, and ozonation (Al-Tohamy et al. 209
 210 2022; Yusuf et al. 2020). These techniques, however, have 210



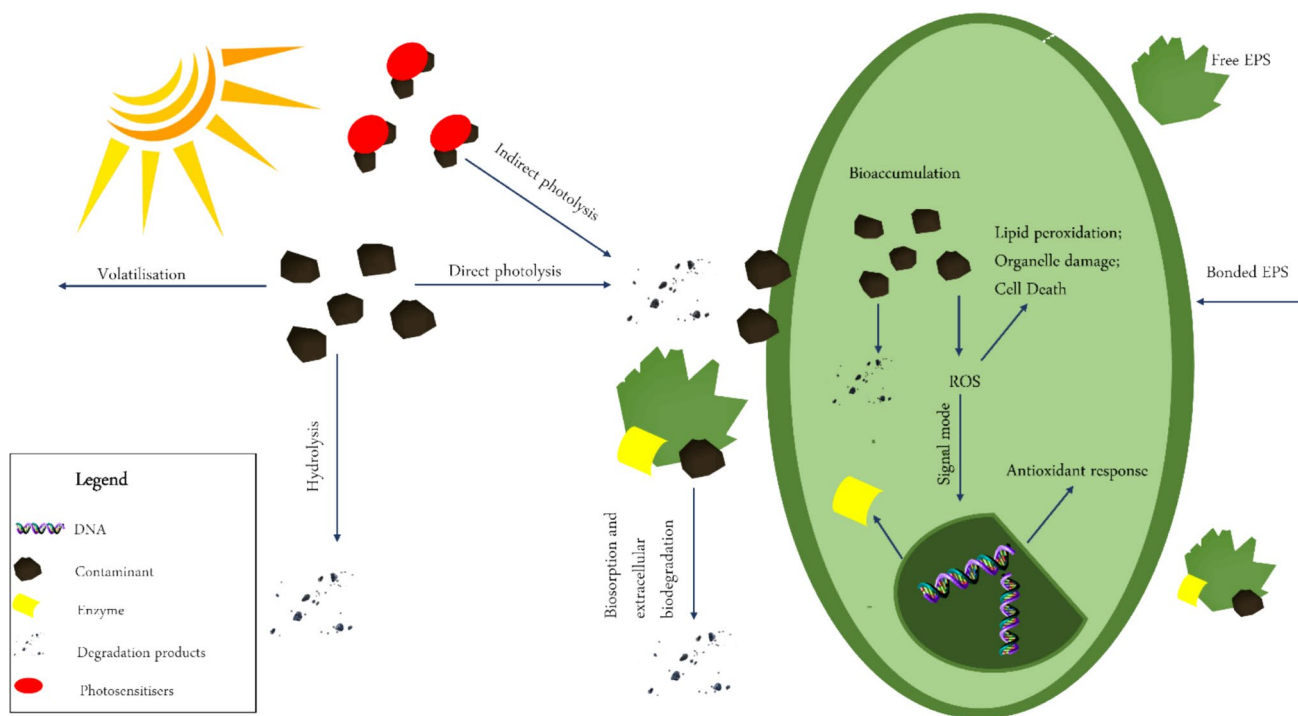


Fig. 4 An overview of metabolic processes that microalgae employ for pollutant removal (adapted from Zhou et al. 2022)

211 several drawbacks including high operating costs, genera-
212 tion of wastes and by-products, and removal inefficiencies.

213 Biological treatment of industrial wastewater using
214 microalgae has gained significant attention as a potential
215 treatment option. Microalgae can metabolize hazardous
216 materials into non-toxic substances, which makes it one
217 of the most promising and sustainable alternatives to con-
218 ventional treatment methods. Microalgal treatments for
219 wastewater are mostly implemented using pond systems or
220 high-rate algal ponds in domestic wastewater in commercial-
221 ized treatment approaches (Wollmann et al. 2019). These
222 are largely grouped as open systems. A major advantage of
223 algal-mediated wastewater treatment is that it targets emerg-
224 ing contaminants with efficiency with low operating costs.
225 This has increased their appeal for industrial, domestic, and
226 agricultural wastewater treatment.

227 Treatments applying algal consortia are typically applied
228 in outdoor conditions however, some industrial effluents
229 require monitored conditions for treatment due to extreme
230 parameters such as acidity, temperature, and nutrient loads
231 of raw effluent. These systems are classified as closed sys-
232 tems. In these cases, implementation of algal treatment has
233 been reported at laboratory and pilot scale, with effluent vol-
234 umes treated ranging from a few liters to just under 1 KL.

235 In high-rate algal ponds (HRAPs), the system consists of
236 symbiotic algae and bacterial growth, within raceway ponds.
237 The benefit of this system is the ability to recover algal bio-
238 mass as well as the ability of the system to remove emerging
239 contaminants such as antibiotics. Other contaminants such
240 as hydrocarbons and surfactants have also been efficiently
241 removed in HRAPs (Matamoros et al. 2015). Quite often the
242 use of the recovered biomass is applied in biofuel produc-
243 tion or other commercially valuable products, with biomass
244 yields often being a key factor in the viability of this treat-
245 ment (Sutherland et al. 2020).

246 A comparison of differently sized HRAPs found that
247 although nutrient removal occurred at large, pilot and mes-
248ocosm size, optimal conditions, and yield were observed
249 at smaller scales. In a similar study, the modification of
250 the retention times of full-scale algal treatments led to an
251 improvement in the final effluent quality as well as the over-
252 all efficiency of the HRAP performance (Sutherland et al.
253 2020).

254 Among the open systems, stabilization ponds are also a
255 common solution, driven by similar factors; the response
256 to emerging contaminants and a need for feasible treatment
257 solutions. Waste stabilization ponds are non-aerated/mixed
258 systems, unlike HRAPs. The system depends on sunlight

Table 1 Studies on industrial wastewater treatment efficiencies of various microalgal species

Algal species	Industry type	Removal efficiencies	Type of WW treatment	Reference(s)
<i>Scenedesmus</i> sp.	Fertilizer wastewater plant effluent	COD–93%; BOD–84%; NH ₄ –93%; NO ₃ –84%, PO ₄ ³⁻ –97%	Flask Bioreactors, under laboratory conditions	Pham et al. (2020)
<i>Scenedesmus</i> sp.	Pulp and paper mill effluent	COD–75%; BOD–82%; NO ₃ N–65%; PO ₄ –P–71.29%	Outdoor open ponds	Usha et al. (2016)
<i>Planktochlorella Nurekis</i>	Pulp and paper mill effluent	Cr–100%; Co–97%; Ni–71%; Zn–74%; As–98%; N–95.6%; P–100%; S–80%; COD–92.8%	Flask Bioreactors, under laboratory conditions	Sasi et al. (2020)
<i>Chlamydomonas reinhardtii</i>	Pulp and paper mill effluent	Cr–100%; Co–47%; Ni–49%; Zn–68%; As–57%; N–85.6%; P–87.5%; S–60%; COD–91.9%	Flask Bioreactors, under laboratory conditions	Sasi et al. (2020)
<i>Oocystis</i> sp.	Fish processing wastewater	COD- 71.1%; P-70%	Flask Bioreactors, under laboratory conditions	Riano et al. (2011)
<i>Chlorella zofingiensis</i>	Piggery wastewater collected from the private farm	TN–82.70%; COD–68.96%	Flask Bioreactors, under laboratory conditions	Zhu et al. (2013)
<i>Pithopora</i> sp.	Thermal wastewater	COD–87.23%; BOD–87.75%	Flask Bioreactors, under laboratory conditions	Murugesan and Dharmotharan (2009)
<i>Botryococcus</i> sp	Laundry effluent	COD–70.68%; BOD–61.11%; TOC–76.66%; TN–72.49%; P–78.23%; N–86.62%	Flask Bioreactors, under laboratory conditions	Gani et al. (2015)
<i>Chlorella pyrenoidosa</i>	Textile effluent	Cr–80%; Cu–85%; Pb–80%; Cd–91%	Flask Bioreactors, under laboratory conditions	Soeprbowati and Hariyati (2017)
<i>Chlorella</i> sp. and <i>Scenedesmus</i> sp.	Textile effluent	TDS–89.1%; BOD–91.5%; COD–91.9%	Flask Bioreactors, under laboratory conditions	Aragaw and Asmare (2018)
<i>Scenedesmus</i> sp.	Tannery effluent	Cr–57%; Cu–79%; Pb–48%; Zn–65%	Flask Bioreactors, under laboratory conditions	Ajayan et al. (2015)
<i>Scenedesmus</i> sp.	Tannery effluent	N–90%; P–99.85%; S–92.68%; BOD–88%	Flask Bioreactors, under laboratory conditions	Ballen-Segura et al. (2016)
<i>Oscillatoria</i> sp. and <i>Wesiellopsis</i> sp.	Dye effluent	Colour removal–75%; COD–95%; Cl–50%; Mg- 90%	Flask Bioreactors, under laboratory conditions	Vijayakumar (2005)
<i>Chlorella conglomerate</i>	Soft drink industrial effluent	BOD–77.0%; COD–77.7%; Mg–66.7%; P–50%	Outdoor open ponds	Sivasubramanian et al. (2012)
<i>Chlorococcum humicola</i>	Soft drink industrial effluent	BOD–86.1%; COD–85.9%; Mg–62%; P–50%	Outdoor open ponds	Sivasubramanian et al. (2012)
<i>Chlorella vulgaris</i>	Brewery effluent	BOD–91.43%; COD–83.11%; TN–83.74%; TP–54.69%	Flask Bioreactors, under laboratory conditions	Choi et al. (2016)
<i>Scenedesmus obliquus</i>	Brewery effluent	COD–57.5%; TN–20.8%	Flask Bioreactors, under laboratory conditions	Mata et al. (2012)
<i>Chlorococcum humicola</i>	Oil drilling industrial effluent	BOD–96.17%; COD–52.19%; Cr–92.57%; Cu–88.17%; NO ₃ –78.04%	Flask Bioreactors, under laboratory conditions	Sivasubramanian and Muthukumar (2012)



259 and microalgae-driven nutrient removal. These systems
260 can be decentralized and are often applied as secondary or
261 tertiary water treatment. The low operating costs (e.g., no
262 need for electricity as these are systems are gravity-fed) of
263 such systems have led to their suitability in low-income or
264 developing countries, especially in tropical and subtropical
265 regions (Kaloudas et al. 2021). Challenges in biomass recovery,
266 hydraulic retention times, temperature, sunlight, as well
267 as contamination (e.g., protozoa) are however factors that
268 affect the optimal functioning of this open system (Arora
269 et al. 2021; Mahapatra et al. 2022).

270 Closed systems comprise examples of photobioreactors,
271 where biomass recovery, volumes, and effluent quality can
272 be closely monitored and optimized. The usage of these systems
273 is beneficial where effluent is industrial or contains
274 high nutrient loads. Floating bioreactors or immobilized
275 algae, either through alginate beads or biofilms, are examples
276 of smaller-scale treatments where algal biomass can
277 be recovered more easily. A few of these technologies have
278 been commercialized globally, such as the revolving algal
279 biofilm for example, as well as a suspended photobioreactor
280 that treats up to 50 000 gallons of wastewater per day (Wol-
281 lmann et al. 2019). The operational and maintenance costs
282 however, of these closed systems are substantially more, and
283 although improved, the challenge in enhancing the recovery
284 of biomass due to size and the usage of flocculation and/or
285 other immobilizing techniques remains immanent.

286 According to Amin et al. (2022), the main objective of
287 wastewater treatment is to reduce nutrient levels and loads
288 to concentrations that are acceptable to the regulating bodies
289 with respect to the final effluent discharged. To achieve
290 this, the application of co-culture methods has also proven
291 effective in the successful treatment of industrial effluent
292 (Qi et al. 2018; Perera et al. 2022). This involves the co-
293 culturing of algal isolates with fungal or bacterial isolates
294 for enhanced treatment.

295 The natural processes associated with microalgae-driven
296 treatment are advantageous from an environmental, potential
297 cost, and infrastructure perspective, as well as a robust
298 response to emerging contaminants. Moreover, the ability
299 to decentralize treatments and modify retention times provides
300 a responsive, efficient solution that services the specific
301 needs of a community or area. However, these benefits
302 are realized when applied on a larger scale open system, as
303 the costs linked to closed, smaller scale treatment on harsh
304 effluent may be at a higher operating cost. In addition, the
305 impacts of climate, temperature, and the need to optimize
306 biomass recovery due to cell size are some of the continued
307 challenges, especially in coupling this with any by-product
308 generation from algae.

Algae-assisted industrial wastewater treatment

309
310

311 The industrial sector is an important consumer of freshwater
312 resources and a significant contributor to waste. In South
313 Africa, industry is one of the largest consumers of water for
314 development (DWS 2017), and globally consumes 22% of
315 water (UN Water 2018). Similarly, the sector represents one
316 of the largest contributors to wastewater generation (van der
317 Merwe et al. 2009; Cloete et al. 2010), with an estimated
318 11.5 million tons of brine per year. Nationally, water consumption
319 and effluent production volumes are highest in the
320 petroleum and paper and pulp industries (Cloete et al. 2010).
321 Treated wastewater from various industry types has different
322 properties, such as colour, pH, TDS, TSS, BOD, COD, total
323 solids (TS), electric conductivity (EC), and heavy metals
324 which are often above the permissible limits.

325 The composition of industry wastewater is highly diverse
326 and generally includes pesticides, dioxins, heavy metals,
327 poly-aromatic hydrocarbons (PAHs), polychlorinated
328 biphenyls (PCBs), phenolic compounds, petrochemicals and
329 microorganisms (Kanu and Achi 2011). These contaminants
330 overall have negative effects on human life and aquatic life.
331 Phosphate nutrients in wastewater are used in algae for the
332 synthesis of nucleic acids, phospholipids, and phosphate
333 esters (Benitez-Nelson 2000). Like phosphorous, nitrogen-
334 based contaminants in wastewater are assimilated for algal
335 growth and development (Kumar et al. 2013; Singh et al.
336 2021).

337 The bioremediation of the most common industrial effluent
338 using algae is discussed in the sections below.

Effluents from the paper and pulp industry

339

340 The pulp and paper industry generates significant amounts of
341 waste from its various processes (Patel et al. 2021; Esmaceli
342 et al. 2023). According to Chaudhry and Paliwal (2018),
343 besides lignin and naturally occurring polymers, the xenobiotic
344 compounds generated by the industry include chlorinated
345 lignin, chlorinated phenol, chlorinated resin acid,
346 dioxins, chlorophenols, phenols, adsorbable organic halogens
347 (AOX), and extractable organic acids, halogens, and
348 metal ions. Several algal species have been reported for their
349 ability to remove contaminants from the wastewater of paper
350 industries. For example, paper and pulp effluent was treated
351 using *Planktochlorella nurekis* and *Chlamydomonas reinhardtii*,
352 and the results reported a nitrate reduction (96%
353 and 86%), phosphate (100% and 88%), COD (92% and 93%)
354 and heavy metals (100% Cr, 97% Co, 77% Ni, 71% Cu, 72%



355 Zn and 88% Cd). Sasi et al. (2020) reported similar results
 356 for the same algal species. Another study reported on the
 357 benefits of mixed algal strains of *Chlorella*, *Chlorococcum*,
 358 *Chlamydomonas*, *Pandorina*, and *Eudorina*, in removing
 359 75% COD, 84% colour, and 80% AOX (Tarlan et al. 2002).

360 Effluents from the petrochemical industry

361 Effluent from the petrochemical industry contains an array
 362 of contaminants including organic solvents, phenolic com-
 363 pounds, aniline, nitrobenzene, and sulfides (Tian et al. 2020;
 364 Cechinel et al. 2016). Microalgae have been reported to be
 365 promising for the treatment of wastewater from the petro-
 366 chemical industry (Madadi et al. 2016; Madadi et al. 2021).
 367 Specifically, the algae strain *Chlorella vulgaris* was found
 368 to be the most effective for the removal of hydrocarbon with
 369 98% efficiency (Al-Hussieny 2020).

370 Phycoremediation has also been reported to be a promis-
 371 ing technique for the tertiary treatment of phenol-polluted
 372 petrochemical effluents through the degradation process
 373 (El-Gendy and Nassan 2021). Eukaryotic microalgae were
 374 reported to degrade phenol to less toxic β -ketoadipate (Das
 375 et al. 2015), while *Ochromonas Danica* was reported to
 376 degrade phenol to cis-2-hydroxypenta-2.4-dienoate (Sem-
 377 ple and Cain 1996).

378 A combination of bacteria and algae has been reported to
 379 be effective in the treatment of phenol-polluted wastewater
 380 (Nassar and El-Gendy 2021). Bacteria provide algae with
 381 carbon dioxide needed for photosynthesis while algae pro-
 382 vide the bacteria with oxygen for their growth (Liang et al.
 383 2013). Light was reported to be essential for the biodegrada-
 384 tion of high concentrations of phenol in algae and bacteria
 385 combinations. A combination of the algal strain *Chlorella sp*
 386 with bacteria resulted in a complete removal of 1200 mg/L
 387 of phenol (Nassar and El-Gendy 2021). A recent study by
 388 Madadi et al. (2021) reported promising results for lipid pro-
 389 duction at the same time as treating petrochemical wastewa-
 390 ter using *Chlorella vulgaris*.

391 Effluents from acid mine drainage

392 The high acidic effluent from acid mine drainage (AMD)
 393 can be harmful to wastewater treatment plants and the eco-
 394 system. Acidic effluent contains high levels of heavy metals,
 395 and hence, there is a need for an economically and eco-
 396 friendly method for removing heavy metals in the effluent.
 397 Algae can absorb or bond with metals based on the heavy
 398 metal biosorption properties, a process that results in a
 399 higher pH that eliminates metals from wastewater. At high

pH, heavy metals are not soluble in water (Monteiro et al. 400
 2012). Algal strains including *Spirulina*, *Chlorella*, *Ana- 401
 baena*, and *Oscillatoria* have been reported for the effective 402
 treatment of AMD wastewater (Shinde et al. 2018). A recent 403
 laboratory-scale study by Santos et al. (2020) reported the 404
 use of *Scenedesmus sp* for successful treatment and removal 405
 of metals from coal mining AMD, while also producing 406
 algal biomass. 407

Effluents from the pharmaceutical industry 408

409 Pharmaceuticals and personal care products are of emerg-
 410 ing concern due to their substantial usage and persistent
 411 nature. The growth of the pharmaceutical industry poses
 412 a threat to wastewater treatment processes as many waste-
 413 water treatment plants are not fully equipped to remove all
 414 pharmaceutical contaminants in wastewater (Goswami et al.
 415 2022). Wastes generated by the pharmaceutical industry are
 416 as diverse as their processes and operations, however gen-
 417 erally contain antibiotics, antiepileptics, psychiatric drugs,
 418 anti-inflammatories, diuretics, analgesics, and β blockers
 419 (Mustafa et al. 2021). Biosorption, bioaccumulation, and
 420 biodegradation have been reported as the main processes to
 421 remove pharmaceutical contaminants in wastewater (Gentili
 422 and Fick 2017; Gojkovic et al. 2019; Sun et al. 2023). Micro-
 423 algal remediation of the effluents from the industry has been
 424 reported. For example, Gentili and Fick (2017) reported 70%
 425 N and 89% P removal by *C. sorokiniana*, whereas Escapa
 426 et al. (2015) reported only 67.8% N and 55.6% P removal by
 427 *Dictyosphaerium*. Similarly, Prathiba et al. (2020) reported
 428 high removal efficiencies in reducing BOD, COD, TSS, and
 429 TDS by *Chlorella* and *Scenedesmus sp*.

Effluents from the food and beverage industry 430

431 The food and beverage industry is generally a large con-
 432 sumer of water and a large producer of wastewater effluent
 433 with high levels of nutrients and organic compounds; and
 434 hence has high levels of COD, volatile acid, total nitrogen,
 435 total phosphorous, and heavy metals. Several food process-
 436 ing companies are unable to meet the regulations imposed
 437 by municipalities for the disposal of effluent to wastewater
 438 treatment plants. Furthermore, the current methods used to
 439 reduce the level of contaminants in effluent from the industry
 440 are often costly, and hence, there is a need for less expensive
 441 wastewater purification methods. Microalgae-based methods
 442 of wastewater treatment from the food and beverage indus-
 443 try have been reported. Daneshvar (2019) observed 76.7%
 444 organic carbon, 92% total N, and 100% P removal by *S.*

445 *quadricauda* from dairy wastewater. Song (2020) reported
 446 90% ammonia, 76% total nitrogen, 95% total phosphate, and
 447 74% chemical oxygen demand removal by *Scenedesmus sp.*
 448 from brewery waste. Mawi et al. (2020) reported nitrate
 449 removal of up to 58.8% and ammonia removal of 72.5%
 450 from wastewater from a shrimp industry by using *Gracilaria*
 451 *edulis* and *G. changii* in a biofilter. The microalgae
 452 *Scenedesmus dimorphus* was able to remove more than 99%
 453 of both N and P from brewery wastewater within one week
 454 (Lutzu et al. 2016).

455 Future prospects of phycoremediation

456 The future prospects of algae treatment and the need for
 457 improvements in phycoremediation to allow for scalability
 458 and sustainable wastewater treatment for industry are sum-
 459 marised in Fig. 5. As technology continues to advance, there
 460 is increasing potential for the development of more efficient
 461 and cost-effective algae cultivation and wastewater treatment
 462 systems. These advancements encompass improvements in
 463 photobioreactors, cultivation techniques, and genetic engi-
 464 neering of algae strains to bolster their pollutant removal
 465 capabilities (Razzak et al. 2023). Industries are exploring
 466 the integration of algal bioremediation into their sustain-
 467 ability initiatives, cultivating algae either on-site or nearby
 468 to treat wastewater before release. This strategic approach
 469 minimizes the environmental footprint and reduces expenses
 470 associated with regulatory compliance (Sousa et al. 2022;
 471 Razzak et al. 2023).

472 Research efforts are anticipated to concentrate on iden-
 473 tifying and engineering novel algal strains with enhanced
 474 pollutant removal capabilities. This may involve selecting
 475 strains that thrive in specific industrial effluents or utilizing
 476 genetic modification to optimize nutrient uptake and pollut-
 477 ant degradation (Abdelfattah et al. 2022; Hassanien et al.
 478 2023). Furthermore, scaling up algal bioremediation systems
 479 to meet the demands of large industrial facilities is a key
 480 focus. This could entail the development of modular systems
 481 that can be easily expanded or replicated to accommodate
 482 varying wastewater volumes and pollutant loads (Goh et al.
 483 2023).

484 Algae-based wastewater treatment also offers inte-
 485 gration with circular economy concepts, where treated
 486 wastewater and algal biomass can be reused or repurposed
 487 within industrial processes. For instance, algal biomass
 488 can be converted into biofuels, biofertilizers, or high-value
 489 products, generating additional revenue streams and reduc-
 490 ing waste (Zabochnicka et al. 2022). Collaboration among

research institutions, industries, government agencies, and
 environmental organizations is essential for driving inno-
 vation and knowledge sharing in the field of algal biore-
 mediation (Mellor et al. 2022).

Moreover, regulatory support and incentives play a
 crucial role in driving the implementation of sustainable
 wastewater treatment practices, including algal bioreme-
 diation, in industries (Abdel-Raouf et al. 2012). Increasing
 public awareness of the benefits of algal bioremediation
 can lead to greater acceptance and demand for these solu-
 tions. Education campaigns and outreach efforts can help
 foster support for algae-based wastewater treatment initia-
 tives (Mellor et al. 2022).

While phycoremediation presents promising technol-
 ogy for industrial wastewater treatment, further research
 is necessary to identify suitable algae for maximum pol-
 lutant removal while maximizing biomass production
 (Ahmad et al. 2022). Optimizing growth conditions for
 microalgae on a large scale is imperative for successful
 industrial wastewater treatment. Additionally, advance-
 ments in microalgal engineering and genetic modification
 hold promise for enhancing bioremediation efficiencies
 and increasing the utility of phycoremediation technology
 (Cabanelas et al. 2016; Fajardo et al. 2020; Kumar et al.,
 2020). By capitalizing on these advancements and promot-
 ing collaborative efforts, phycoremediation can emerge as
 a vital tool for achieving environmental sustainability in
 wastewater treatment practices.

Challenges involved in phycoremediation

Phycoremediation is undoubtedly a beneficial strategy for
 remediating the environment and improving the quality
 of industrial wastewater. It has been proven to be an eco-
 friendly and sustainable bioremediation solution. How-
 ever, like any innovation, this technology has some chal-
 lenges that may hinder its optimal operation and efficiency.

Since phycoremediation uses microalgae to remove pol-
 lutants from industrial wastewater, the selection of micro-
 algae is therefore important to ensure maximum efficiency.
 As mentioned in the previous sections, species such as
Scenedesmus sp., *Chlorella vulgaris*, and *Oscillatoria*
limos, have been used successfully to remediate industrial
 wastewater effluent. However, the data also shows that
 the efficiency of microalgae in phycoremediation may be
 species-specific as well as industry wastewater-specific.
 For instance, the use of *Scenedesmus sp.* was effective
 for nutrient removal in tannery wastewater (Ajayan et al.



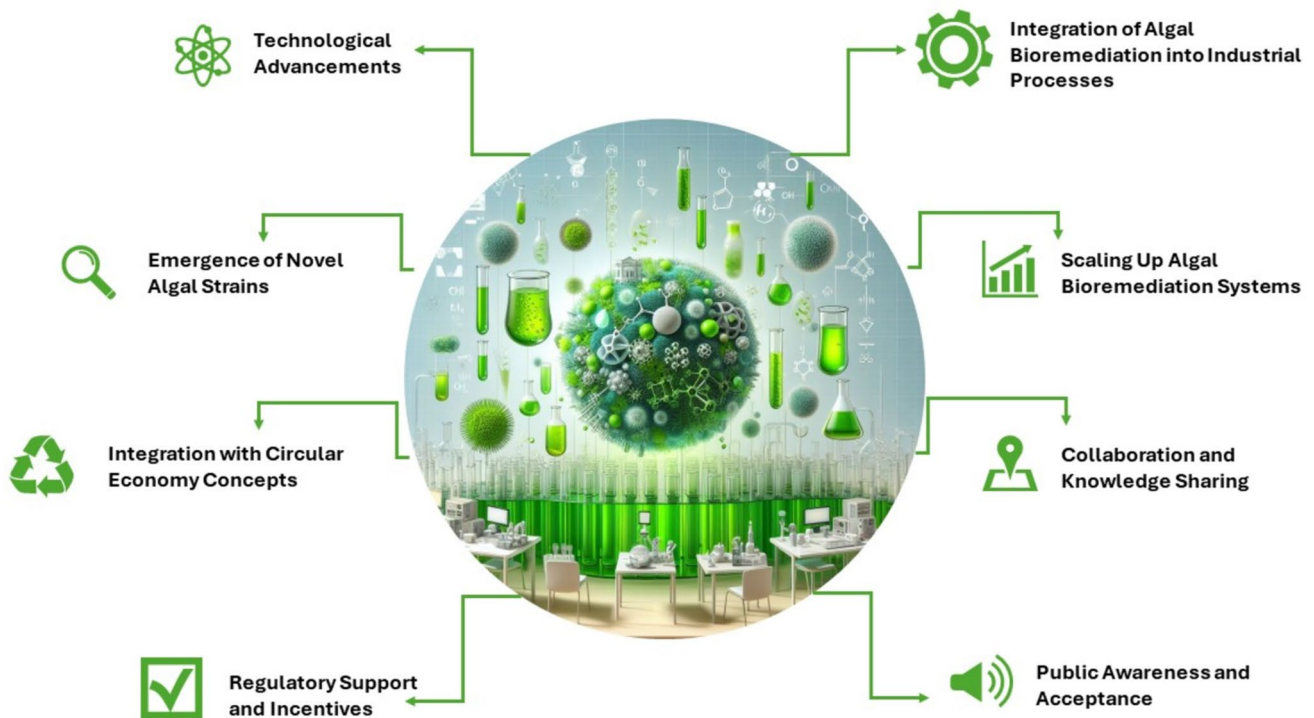


Fig. 5 Key aspects impacting algae-based wastewater treatment in future

2015), but not as efficient for pulp and paper mill industrial wastewater (Amit and Ghosh 2018). However, species such as *Chlorella sp.* are generalists; their efficiency was reported to be high in textile wastewater (*Chlorella pyrenoidosa*), rolling mill wastewater, and tannery wastewater (*Chlorella vulgaris*) (Atoku et al. 2021; Nagabalaaji et al. 2019; Pathak et al. 2014), with the efficiency dependent of target pollutants. Therefore, microalgal culturing should be undertaken in different wastewater types to determine the most efficient species to utilize. To this end, the use of consortia has been recommended to ensure phycoremediation efficiency. This is due to multiple species complementing each other's metabolic processes thus leading to successful bioremediation (Rambabu et al. 2023). Moreover, the inefficiencies of one species due to effluent type are supplemented by the species active in that particular media. The microalgae consortia can be further supplemented by incorporating specific species of bacteria. Algae-bacteria consortia have been shown to increase the growth rate of microalgae and improve their efficiency in pollutant removal (Mujtaba et al. 2016). For the success of this consortia, adequate concentrations of CO₂, light

intensity, and nutrients need to be optimized to ensure the successful growth of both the microalgae and bacteria (Chia et al. 2021).

In addition to species selection, microalgal cultures must be successfully reared. Climatic conditions have been reported to affect algal growth and efficiency. For instance, low temperatures have been shown to prohibit biological processes and reduce the removal of contaminants (Kumar et al. 2021). As a result, there is little evidence of successful phycoremediation in colder climates. Algae rearing presents additional challenges in addition to climatic conditions. For example, algae reared in open systems has been reported to be vulnerable to contamination (Hejna et al. 2022; Kumar et al. 2021). This leads to negative effects on the successful cultivation of microalgae. In closed systems such as bioreactors, microalgal biofilms form on the inner walls of photobioreactors limiting light penetration and reducing the supply of CO₂ to the culture medium and the build-up of O₂ generated during photosynthesis (Rambabu et al. 2023).

Another major constraint is that some pollutants have been reported to inhibit the growth of microalgae. For instance, Atoku et al. (2021) utilized *Nostoc commune*,

581 *Oscillatoria limos*, and *Chlorella vulgaris* to remove heavy
 582 metals (Cu, Fe, Ni, Pb, Cd, Zn) from raw wastewater from
 583 the rolling mill industry. The phycoremediation process
 584 was able to address the physicochemical parameters such
 585 as pH, DO, and BOD. Chloride, nitrates, and sulphates
 586 were reduced by 59–63%, 93–99%, and 60–70%, respec-
 587 tively (Atoku et al. 2021). Although there was a significant
 588 removal of Pb (71–90%), Cd (49–54%), Cu (9–10%), Fe
 589 (27–30%), Ni (31–38%), and Zn (24–27%), the average
 590 concentrations of heavy metals in the final effluent did not
 591 meet the required local standard for the final effluent (Atoku
 592 et al. 2021). The removal efficiency of heavy metals from
 593 raw wet market wastewater produced similar results. The
 594 use of *Scenedesmus spp.* had low removal of Cd (47.2%),
 595 Zn (51.75%), and Fe (62.57%) (Apandi et al. 2018). Only
 596 Cr had a significant removal of 89.2% (Apandi et al. 2018).
 597 In contrast, there was a higher removal efficiency of heavy
 598 metals in the diluted (50%) wet market wastewater (Apandi
 599 et al. 2018). This is due to the toxic nature of the industrial
 600 wastewater to microalgae resulting from high pollutant load
 601 (Saranya and Shanthakumar 2019). The authors reported
 602 that textile wastewater- inhibited the growth of microalgae
 603 biomass to below control levels. Thus, pollutants' toxicity
 604 affects the optimal growth of the algae and their efficiency.
 605 To this end, phycoremediation can only be used as a supple-
 606 mentary tertiary treatment, and pre-treatment of the waste-
 607 water is required to ensure maximum efficiency.

608 The scale-up challenge has also been reported. Phycore-
 609 mediation on a smaller scale (laboratory setting and small
 610 domestic ponds) has been piloted successfully (Angelaal-
 611 incy et al. 2023; Cavieres et al. 2021; Sivasubramanian et al.
 612 2009; Oberholster et al., 2019). However, the scale-up has
 613 been shown to reduce the efficiency of the process of pollut-
 614 ant removal (Kumar et al. 2021).

615 Perhaps the biggest challenge in the use of phycoremedi-
 616 ation at a large scale is the harvesting of algae biomass
 617 (Hejna et al. 2022). Microalgae cells are small and have
 618 a density close to that of water (1.08–1.13 g/cm³), have a
 619 strong surface charge, and are highly mobile (Leong et al.
 620 2021). The harvesting process is costly; 30% of total pro-
 621 duction expenses are incurred for harvesting algae biomass
 622 (Kumar et al. 2021; Pacheco et al. 2015). To this end, several
 623 methods have been investigated in an attempt to recover the
 624 microalgae biomass. Such methods include centrifugation,
 625 flocculation, filtration, flotation, sedimentation, and electro-
 626 coagulation (Grima et al. 2003; Morais Junior et al. 2020).
 627 Centrifugation is effective, especially for the recovery of
 628 high-nutrient biomass useful in resource recovery (Najjar &
 629 Abu-Shamleh 2020). However, centrifugal harvesting only
 630 works well for cells with high settleability (e.g. *Scenedesmus*

631 *spp.*) and with a high rotation rate of the centrifuge (a low
 632 rate has been reported to reduce harvest efficiency by up to
 633 55% (Dassey and Theegala 2013)). Furthermore, centrifuga-
 634 tion is economically unviable since the process is energy-
 635 intensive and the financial cost is too high (Vandamme et al.
 636 2013). On the other hand, while the sedimentation process
 637 is efficient, the recovered biomass is generally too dilute
 638 and compromises resource recovery (Mathimani & Mallick
 639 2018). Due to the use of membrane technology in the filtra-
 640 tion process, this recovery method is only viable on a small
 641 scale (Salim et al. 2011; Zhu et al. 2018). In the floccula-
 642 tion process, large concentrations of inorganic flocculants
 643 are required which leads to high production of sludge, is
 644 only effective for some algal species, and the final prod-
 645 uct is highly contaminated due to the addition of inorganic
 646 salts (Chen et al. 2011; Christenson & Sims 2011; Singh &
 647 Patidar 2018). Overall, the harvesting method used must be
 648 able to process large volumes of algal biomass, at a large
 649 scale, while keeping the cost at a minimum. Once the har-
 650 vesting has been optimized, the successful and sustainable
 651 recovery of resources (by-products) can be honed.

652 Conclusion

653 Phycoremediation presents a viable alternative to conven-
 654 tional wastewater treatment processes. The technology has
 655 been primarily recommended due to its resource recovery
 656 capabilities and has the potential to produce biodiesel,
 657 biogas, and biofertilizer (De Mendonça et al. 2021). At
 658 the laboratory scale, the production of biofuel has already
 659 been evidenced. However, without successful and opti-
 660 mal biomass harvesting, this remains a challenge. So far,
 661 the production of biodiesel and biogas is energy-intensive. AQ1
 662 This is in addition to challenges related to biomass recov-
 663 ery. Furthermore, the sustainable and economically viable
 664 recovery of primary and secondary metabolites needs further
 665 research. Moreover, the low biomass yield and complicated
 666 contaminants removal (targeting fit-for-purpose utilization)
 667 compromise the full circular economy implementation of the
 668 phycoremediation process. The benefits of phycoremedia-
 669 tion have been highlighted in this review. Phycoremedia-
 670 tion presents a green approach for contaminant removal to
 671 support environmental protection and a long-term remedia-
 672 tion option. As evident, removal efficiencies are not only
 673 dependent on the algae use but also on the type of waste-
 674 water. However, much more investigation is needed to support
 675 the current research. Future research should therefore focus
 676 on the selection of effective algal species, cultivation meth-
 677 ods, and a combination of other physicochemical variations.



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681 **Declarations**

682 **Conflict of interest** The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

685 **Ethical approval** This article does not contain any studies with human participants or animals performed by any of the authors.

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