1 2 3	A comparative assessment of water use by <i>Acacia longifolia</i> invasions occurring on hillslopes and riparian zones in the Cape Agulhas region of South Africa.	
4	Y.P Mkunyana <sup>1*</sup> D. Mazvimavi <sup>1</sup> S. Dzikiti <sup>2</sup> Z. Ntshidi <sup>1, 2</sup>	
5 6	<sup>1</sup> Institute for Water Studies & Department of Earth Sciences, University of the Western Cape, Bellville, South Africa. Emails: <u>3137866@myuwc.ac.za*; dmazvimavi@uwc.ac.za</u>	
7 8	<sup>2</sup> Council for Scientific and Industrial Research (CSIR), Natural Resources and Environment, Stellenbosch, South Africa. Emails: <u>sdzikiti@csir.co.za; zntshidi@csir.ac.za</u>	
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10	Abstract	Commented [S1]: You use the phrases "transpiration rates", "
11	The detrimental impacts of invasive alien plants on ecosystems and water resources have	acacia longifolia" and "invasive alien plants" too repetitively which makes the abstract not so lekker.
12	raised concerns in dry_countries like South Africa where the_average precipitation is	
13	approximately 500 mm/year, which is below the world average of around 860 mm/year.	
14	Several studies have examined the effects of invasive alien plants such as the Australian	
15	Acacias on theSouth Africa's water resources. However, no There has however not been a	
16	study has investigatquantifingied the differences in water use-rates between hillslope and	
17	riparian Acacia longifolia_invasions in South Africa. A.cacia longifolia is one of the	Formatted: Font: Not Italic
18	problematicaggressive invasive alien plantsder species in South Africa and yet- Hhillslopes	
19	contribute substantially to runoff generation, and thereforegeneration. Therefore, the	
20	encroachment of-hillslopes by invasive alien plants occurring on these areas have the can	
21	potential to reduce potentially reduce runoff, thereby and thus adversely affecting the available	Formatted: Strikethrough
22	water resources at the catchment scaledownstream. This paper aims to:; 1) determincompare	
23	transpiration rates of by A. longifolia growing on hillslopes and along riparian areas; 2)	
24	establish the key drivers for transpiration rates water use by this species $a_{\overline{a}}$ and 3) estimate the	
25	hydrological impacts of the invasions at the catchment scale in the Heuningnes catchment	
26	which is in the Western Cape Province of South Africa. The water use rates ranspiration by	
27	the treesA. longifolia wasere measurdetermined using the hHeat ratioPulse Velocity (HPV)	Formatted: Font: Not Italic
28	sap flow $\underline{method}_{\underline{technique}}$ . An automatic weather stations and soil moisture sensors were	
29	used to monitor weather and soil water content variations at each site. The results showed	
30	that, at the stand level the riparian A. longifolia transpired two times more water ( $\simeq$ 596	
31	mm/yr) than on the hillslope (~242 mm/yr). During years with above average above the	
32	average-rainfall-above the average, the the water use rates by A. longifoliathe invasions was	Formatted: Font: Not Italic
33	estimated to be $\geq$ 579 mm/yr on the hillslope and could be as much as $\geq$ 1 348 mm/yr at the	
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riparian site. Thus, the hypothesis that riparian trees use more water than invasions on nonriparian areas was accepted in this study. <u>At-theAt the</u> catchment scale, the estimated water use by <u>the invasionsve alien plants</u> was 20.5 Mm<sup>3</sup>/year. Clearing of all the invasions in the study catchment is would likely to-make 17 Mm<sup>3</sup>/year of water available. Hence <u>the clearing</u> of *A. longifolia* along riparian corridors should be prioritised in the riparian areas as this will

6 lead to water savings at the catchment scale.

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9 Most parts of sub-Saharan Africa are semi-arid to arid with actual evapotranspiration dominating the terrestrial hydrological cycle. Water resources assessments such as the 10 WR2012 for South Africa (Bailey and Pitman, 2017) show that the mean annual runoff is 11 12 generally less than 10% of the mean annual rainfall. Thus, evapotranspiration is greater than 90% of the annual rainfall in most parts of South Africa and elsewhere in the arid regions. 13 14 Several studies in South Africa have demonstrated that land use and land cover changes often 15 increase evapotranspiration rates which adversely affect the available water resources in a region that is already water scarce (Scott et al., 2006; Bullock and Acreman, 2003). The 16 17 invasion of indigenous vegetation by Invasive Alien Plants (IAPs) has been shown to reduce the available water resources due to their higher water use rates. Studies by Enright (2000), 18 19 Le Maitre et al. (2002), Chamier et al. (2012), Dzikiti et al. (2013), and Meijninger and Jarmain (2014) showed that the Australian Acacia, Eucalyptus and Pinus genera have the 20 21 largest impact on South Africa's water resources. The increase in evapotranspiration rates 22 reduce stream flows (Prinsloo and Scott, 1999), and lowers groundwater levels (Scott et al., 23 2008; Dzikiti et al., 2013b;). Besides their hydrological impacts, invasive alien plants threaten the biological diversity by eroding gene pools, outcompeting the indigenous species 24 25 which consequently resulting in the extinction of endemic species, especially in freshwater ecosystems (Bonanno, 2016). They also occupy invade and occupy grazing land thereby 26 threatening the livelihoods of farmers (Ndhlovu et al., 2011). 27

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The water use rates of oOne of the problematic invasive alien plants (IAP) in South Africa, is *Acacia longifolia*, with water use rates that have never been investigated. *Acacia longifolia*,
which is commonly known as the "long-leaved wattle", is an evergreen shrub/tree with bright

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yellow flowers that sprout from August to October in South Africa. This plant can grow 1 2 between 2 to 8 meters in height (Morais et al., 2015). The species was primarily introduced in 3 South Africa around 1864 for ornamental purposes and for sand dune stabilization 4 (Richardson et al., 1997). However, this species has now spread beyond the initially planted 5 area and has becaome highly invasive. Acacia longifolia is highly problematic in the wetter parts of the Western Cape, Eastern Cape, Kwa-Zulu Natal and parts of Mpumalanga Province 6 7 (Henderson, 2001). Consequently, A. longifolia is considered as a-Category 1b invader, according to the South African National Environment Management: Biodiversity Act (DEA, 8 9 2014). This means that 1) importing into the country, 2) growing, breeding or any other form 10 of propagation, and 3) selling or trading in any specimen of the species are prohibited to reduce the spread and impacts of these invasions. 11

Invasive alien plants occurring within riparian zones tend to have access to multiple sources 12 of water resulting in very high transpiration rates (Le Maitre et al., 2000; 2004). Invasive 13 alien plants are also spreading into hillslopes in mountainous areas of countries like South 14 15 Africa. The mountain areas are regarded globally as strategic water source areas or water towers that require high levels of conservation to ensure sustainable water supply (Messerli et 16 al., 2004). In South Africa, 50% of the runoff is generated from 8% of the land area 17 comprising mostly mountain catchments (Nel, 2013). The spread of IAPs onto hillslopes is 18 19 therefore a major threat to the availability of water resources. However, few studies have compared the water use rates of IAPs growing along riparian zones and on hillslopes. 20

21 Catchments in the Cape Agulhas region are experiencing rapid expansion of areas affected by IAPs along both riparian zones and hillslopes (Kotze et al., 2010). These catchments have 22 headwaters in mountains and low gradient coastal lowlands. Organizations involved with 23 alien plant clearing, such as South Africa's Working for Water Programme, prioritise their 24 25 operations based on the assumption that riparian invasions use twice as much water as those growing in non-riparian settings. Detailed scientific evidence based on actual measurements 26 of plant water use and its drivers are needed for accurate decision making and for resource 27 28 allocation in alien plant clearing programmes. The catchment scale impacts of water use by IAPs on hillslope and riparian zones are also unknown. The objectives of this study were to; 29 1) quantify and compare the transpiration rates of A. longifolia growing on hillslopes and 30 riparian zones, 2) investigate and document key drivers of water use by invasions-IAPs in the 31 different topographical settings, and 3) estimate the hydrological impacts of the invasions at 32 33 the catchment scale.

# 2 4<u>-2.</u> Materials and methods

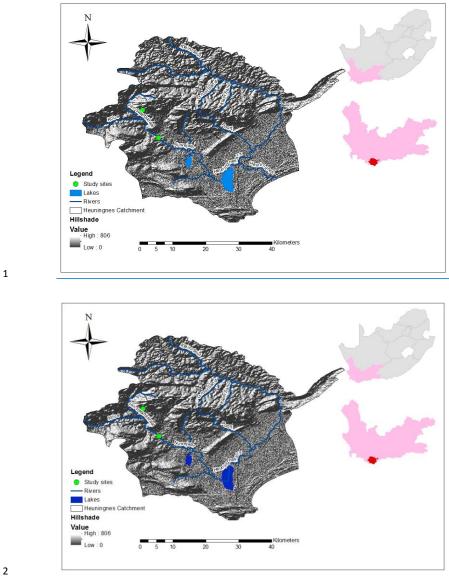
This study was undertaken in the Nuwejaars Catchment from June 2016 to June 2017 which
enabled coveringduring the wet season (June – September 2016) and the dry season
(November 2016 – March 2017).

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# 7 2.1 Study area

8 The Nuwejaars Catchment covers an area of 740 km<sup>2</sup> in the Cape Agulhas region, which is the southern-most part of Africa- (Figure 1). Mountainous areas cover the north-west, 9 10 northern, and northern eastern parts of the catchment with altitude varying ranging from 200 11 to 600 m above sea level. Altitude decreases sharply to less than 100 m after about 10 km along the main river, followed by lowlands that are 10 - 60 m above sea level. Numerous 12 13 wetlands in the form of pans, floodplain wetlands, and lakes occur within the lowland area. The Nuwejaars River passes through a floodplain wetland with the width reaching up to 0.8 14 km. This river discharges into a lake, Soetendalsvlei, which is 3 km wide and 8 km long. 15 Soetendalsvlei discharges into the Heuningnes River that flows into the Indian Ocean. 16



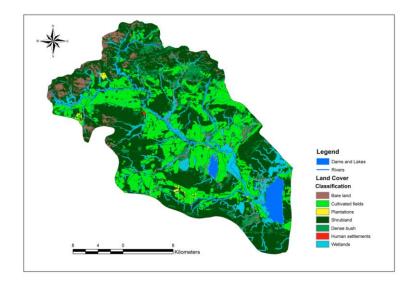


The study area has a Mediterranean climate characterised by wet winters from mid-May to
late August, and dry summers, November to March. Annual rainfall varies from 400 to 600

6 mm/yr, with a mean annual of 650 mm/yr on the hills that form the northern watershed

(Herdien et al., 1995; Kraaij et al., 2009). Maximum temperatures (27 °C) occur in January 1 and the minimum temperatures (8 °C) occur in July and August (Hanekom, Russell and 2 3 Randall, 2009; Herdien et al., 2010; Hoekstra and Waller, 2014). Most streams draining from 4 the mountainous headwaters are perennial due to springs that maintain their flows throughout 5 the year. However, the main river, Nuwejaars River, is non-perennial and mostly flows from May to December, and then dries up for the rest of the year. This river has numerous pools 6 7 that have water throughout the year. The water table is quite shallow, generally less than 3.0 m in the riparian zone and floodplain of the Nuwejaars River. On the hillslopes the depth to 8 9 the water table is variable, 3 - 10 m.

The indigenous vegetation is mainly fynbos, which is a sclerophyllous scrub dominated by 10 species of the Proteaceae, Ericaceae and Restionaceae that are typical of the Cape Floral 11 region (Scott, 1999; Mucina and Rutherford, 2006). In the floodplains, sedges, reeds, restios 12 13 and grass on drier land are dominant. About 41% of the catchment is under agriculture involving commercial dryland crop production (wheat, barley, etc.), sheep and cattle 14 production, irrigated vine yards and orchards, and commercial forestry (Herdien et al., 2010). 15 According to the National Invasive Alien Plant Survey done by Kotze et al. (2010), the 16 invasive alien plants affected\_cover\_265.1 km<sup>2</sup> or 35% of the Nuwejaars Catchment. The 17 dense bushes in the catchments are dominated-mostly formed by invasive alien plantsIAPs 18 (Figure 2). IAPs occur in riparian zones and all the hillslopes in this catchment, and have on 19 20 some locations often form dense and impenetrable stands. Landowners formed a Nuwejaars Wetland Special Management Area forum that aims to improve biodiversity of water-related 21 ecosystems. Coordinated clearing of invasive alien plants is one of the major activities they 22 are undertaking. This complements the Working for Water Programme. 23



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Figure 2: The land cover map of the Nuwejaars catchment

# 4 2.2 Study sites

The study required two sites, a hillslope and a riparian site with actively growing stands of 5 Acacia longifolia. Several sites were evaluated with the assistance of the coordinator of the 6 7 IAP clearing programme run by the Nuwejaars Wetland Special Management Area forum. 8 Since all the land in this catchment is privately owned, permission was sought from the respective landowner to install and regularly monitor the sites. The hillslope site is-was 9 located on the northern part of the catchment and on the southern slopes of the Bredasdorp 10 Hills. The area is locally known as Spanjaardskloof (34° 31' 45.7" S, 19° 45' 9.2" E), and 11 the selected site is-was at an altitude of 125 m and 1.7 km from the hilltop which has an 12 altitude of 430 m. Baboons which can damage monitoring equipment used in this study, This 13 site was selected to minimise damage to monitoring equipment from baboons that frequently 14 15 roam locations withat altitude ranging from 200 to 450 m on this hill. Therefore this hillslope site below 200 m altitude was selected in order to minimize damage of equipment by 16 17 baboons.-Also, tThe presence of suitable trees sizes for transpiration monitoring also influenced the exact site chosen on the hillslope. The riparian site was located less than 20 m 18 from the Nuwejaars River channel which had a 0.8 km wide floodplain in-at the Zoetendals 19

Farm (34° 36' 15.8" S, 19° 47' 46.7" E). This floodplain is locally referred to as the 1 Moddervlei, and has an elevation of 25 m above sea level. A. longifolia occurred as a dense 2 3 stand 120 m wide along the right bank of the Nuwejaars River. The hillslope and riparian sites were 9.5 km apart. Soils at the riparian site varied from deep sandy alluvial soils close to 4 the river channel to dark red clayey loam soils further from the stream channel. At the 5 hillslope site, the soils were shallow loamy soils dominated by boulders and rocks. 6 7 The invasions at the riparian site comprised tall trees with thick stems and an average height between, 7, 0 - 10 m. There was dense undergrowth which that was a mixture of A. longifolia 8 9 trees and the indigenous shrubby tree, Kiggelaria africana (wild peach) trees. Most of the undergrowth was in a poor state and dying either due to the drought during which persisted 10 through the 2016/17 season or due to lack of light as a result of shading by the taller trees. 11 The root system of the riparian trees was fairly shallow, approximately less than 80 cm deep, 12 presumably due to soil water abundance during wet years. The stand characteristics at the 13 14 hillslope site were somewhat different from those at the riparian site, which -reflecting 15 differences in the influenced the growing conditions. Trees at the hillslope site had thinner stems, and they were shorter with the height varying from 3.0 - 6.0 m. The forest floor was 16 mostly clear of vegetation although dense grass occurred in open spaces. The trees had 17 shallow root systems.

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#### 20 2.3 Climate data

21 The microclimate at each site was monitored using an automatic weather station measuring 22 solar irradiance, ambient temperature, relative humidity, wind speed, wind direction, and 23 rainfall. The sensors were installed 2.0 m above the ground and the stations were on a short grass surface. This data was used to calculate the reference evapotranspiration (ETo) 24 according to the FAO 56 guidelines (Allen et al., 1998). All the sensors were connected to a 25 26 data logger that was programmed to store-record data at 15 minutes intervals throughout the 27 duration of the study. Solar panels were used to charge batteries that powered the stations. The hillslope weather station was located at 81 m altitude approximately 1.5 km from the 28 invaded hillslope site, this station was installed for an on-going monitoring project in the 29 30 catchment. The riparian weather station was located 90 m from the invaded site at an altitude

of 25 m above sea level. 31

# 2 2.4 Transpiration, growth, and soil water measurements

Water use rates by Acacia longifolia trees at each site were determined from sap flow 3 measurements using the heat ratio method of the Heat Pulse Velocity (HPV) sap flow 4 technique (Burgess et al., 2001). At the hillslope site, data were collected from three 5 neighbouring trees with stem diameters of 15, 29, and 40 cm at breast height, and 6 7 representative of the stem size distribution at this site. Similarly, three trees with stem diameters of 19, 53, and 58 cm were instrumented at the riparian site. The HPV method has 8 9 been found to be appropriate and cost-effective for transpiration measurement of various tree species (Clulow et al., 2013; Dzikiti et al., 2013, 2016, 2017; Everson et al., 2014; Ntshidi et 10 al., 2018; Scott-Shaw et al., 2017). A metal template with three holes spaced 5 mm apart was 11 used to drill the holes in the stems to minimize probe misalignment. The HPV system 12 13 comprised heaters implanted into the stems and connected to custom-made relay control modules which controlled the heat application. Two T-type thermocouple pairs which 14 measure the sapwood temperature were installed ~0.5 cm above and below each heater probe. 15 16 The thermocouples were connected to multiplexers (Model: AM16/32B Campbell Scientific, Logan UT, USA), which were in turn connected to CR1000 data loggers. Four sets of sensors 17 were installed in the four cardinal directions around the stem on each of the six trees. The 18 sensors were inserted at different depths into the sapwood to account for the radial variation 19 20 in sap velocity (Wullschleger and King, 2000).

Each system was powered using one 105Ah battery. Methylene blue dye was injected in the 21 22 stems to determine the extent of the active xylem vessels (sapwood depth) into which the thermocouples were inserted into. The average bark thickness was about 3 and 5 mm whilst 23 the wood density averaged 0.7 g/cm<sup>3</sup>, and the moisture fraction was 55 % and 63% on the 24 hillslope and riparian site, respectively. The thermocouples were inserted into the sapwood at 25 depths ranging from 12 - 40 mm on the hillslope, and 10 - 50 mm at the riparian site. The 26 27 volumes of water transpired by the instrumented trees were estimated by converting the HPV 28 signals into sap velocities. This was done by correcting for moisture fraction of the wood, wood density, and wounding due to sensor implantation (Swanson and Whitfield, 1981). 29 Total sap flow of the individual trees was calculated by summing the product of the mean sap 30 velocity and sap wood areas corresponding to the sampled areas (Dzikiti et al., 2013; 2016). 31

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1 Tree density at each site was estimated by counting the number of trees in four 10 m x 10 m quadrants. The number of trees per quadrant was then used to estimate the average density of trees per hectare. The stem size distributions were estimated by measuring diameters of forty randomly selected trees at breast height at each site. This information was used to establish stem size classes for each site. Stand level transpiration (*T* in mm/day) was then determined as the weighted sum of transpiration by trees in the respective stem size class as:

$$7 T = \sum SAI_i \times U_i (1)$$

8 where <u>Where</u>  $U_i$  is the average sap flux density in each size class, and  $SAI_i$  is the stand 9 sapwood area index (SAI: m<sup>2</sup> of sapwood per m<sup>2</sup> of ground area).

To estimate the extent of the transpiring leaf area, the leaf area index (LAI) was measured
once in July (wet) and November (dry) using a leaf area meter (Model: LAI 2000, LICOR,
Inc., Lincoln NE, USA). These data were collected on 5 transects at each site at sunset when
diffuse radiation conditions prevailed and the leaves approximated black bodies. Each
transect was about 200 to 300 m long.

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# 16 2.5 Monitoring soil water content

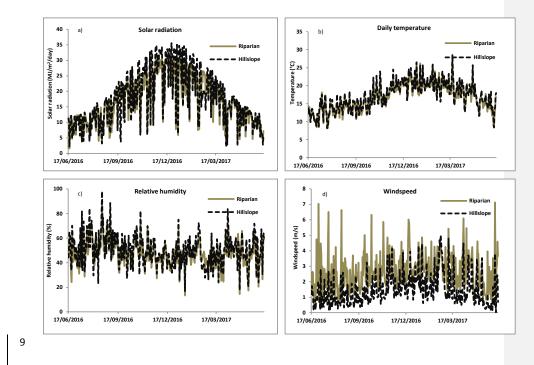
To investigate the influence of the soil moisture variations on tree water use, soil water content was measured in the root zone of the trees at 10, 30, 50, 70, and 90 cm depth at the hillslope and riparian site. Soil water content monitoring was done using 5TE decagon sensors (Decagon Devices, Inc., NE, USA) connected to a data logger (Em50) and programmed to collect data at hourly intervals throughout the study at each site.

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# 23 **5.3.** Results and Discussion

24 3.1 Microclimate and soil water content

There were no major differences in the weather elements between the two sites except for 1 wind speed (Figure 3, Table 1). During the From October 2016 to February 2017, the hillslope 2 3 site received slightly higher radiation, often  $1 - 4 \text{ MJ/m}^2/\text{day}$  greater than the riparian site. The riparian site had generally higher wind speed than the hillslope site. The riparian site is 4 5 located at the centre of a 0.8 km wide U-shaped valley with no windbreaks, and therefore winds moving in a SE-NW direction tend to be funnelled through this valley., hence the high 6 7 wind speed. These conditions resulted in differences in the atmospheric evaporative demand 8 between the two sites.



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 June – August 20	16	November 2016 – February 2017		
Riparian	Hillslope	Riparian	Hillslope	
Min Max	Min Max	Min Max	Min Max	

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Solar Radiation, MJ/m <sup>2</sup> /day	1.6	17.3	3.2	18.7	5.9	31.0	6.0	35.5
Daily temperature (°C)	7.5	18.1	8.1	18.5	14.1	26.3	14.5	26.5
Relative humidity (%)	57.6	92.7	61.0	99.0	50.5	79.2	54.8	86.0
Wind Speed (m/s)	0.3	7.1	0.0	3.8	1.6	6.0	0.6	5.0
Eto (mm/day)	0.8	4.8	0.7	5.3	2.1	10.5	1.8	9.1
Total Rainfall (mm)		277.9		211.0		58.9		111.4
*Total winfall shows the sum of the winfall received and not maximum								

1 \*Total rainfall shows the sum of the rainfall received and not maximum

2 The reference evapotranspiration (ETo) estimated using the Penman-Monteith Equation with

3 an albedo of 0.23 appropriate for deciduous trees was 1243 mm/year and 1306 mm/year at

4 the hillslope and riparian site, respectively. The 5% difference in the reference

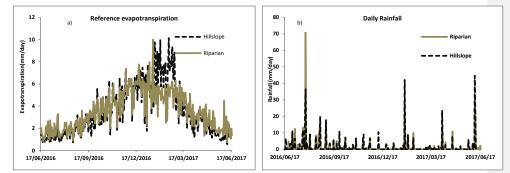
5 evapotranspiration between the two sites is due to higher wind speeds at the riparian site. The

6 daily evapotranspiration rates were similar throughout the monitoring period ranging from

7 0.8 to 10.5 mm/day at the riparian site compared to 0.7 to 9.1 mm/day at the hillslope site.

8 February 2017 was the exception when the hillslope site had generally higher ETo rates than

9 the riparian site (Figure 4a).



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Figure 4: Daily evapotranspiration rates and rainfall at the riparian and hillslopes during
 the<u>from</u> June 2016 – June 2017-period.

Daily rainfall was generally less than 10 mm/day and only exceeded 20 mm/day on 45

occasions during the study period (Figure 4b). The hillslope site received 497 mm/year of

rainfall while the riparian site had 438 mm/year. These rainfall totals are rather low, and are

consistent with the drought that occurred in the entire Western Cape Province during

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1 2016/17. The study period was relatively drier in comparison to long-term average rainfall of

- 2 629 mm/year in 2015 (Mazvimavi et al., 2018).
- 3

# 4 The hillslope and riparian study sites had similar soil characteristics (Table 2). At both sites

5 sandy clay loam occurred from 0 to 30 cm depth.

6 Table 2 Soil textures at various depths on the hillslope and riparian study sites

Depths (cm)	Hillslope	Riparian	
0 - 30	Sandy clay loam	Sandy clay loam	
31 - 60	Sandy loam	Sandy clay loam	
61 - 90	Loamy fine sand	Loamy fine sand	

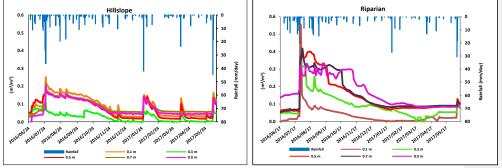
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The Soil Water Content (SWC) in the root zone at both sites showed clear seasonal 8 variations. The soil water content at both sites responded to major rainfall events. On 26 July 9 2016 the riparian site received 72 mm/day of rainfall compared to 36 mm/day at the hillslope 10 site (Figure 5). The water content of the top soil (0.1 m) increased to 0.55 and 0.25 m<sup>3</sup>/m<sup>3</sup> at 11 the riparian and hillslopes sites respectively as ain response to this rainfall event. The soil 12 13 water content at 0.9 m depth increased to  $0.2 - 0.3 \text{ m}^3/\text{m}^3$  at both sites. As from July 2016, 14 soil water content generally decreased at both sites and was  $0.03 - 0.09 \text{ m}^3/\text{m}^3$  by December 2016 (Figure 5). The 42 mm/day of rainfall received at the hillslope site on 27 January 2017 15 caused some increase of soil water content to  $0.10 - 0.20 \text{ m}^3/\text{m}^3$  at this site. The riparian site 16 received 27 mm/day of rainfall on the same day which caused no significant changes in soil 17 water content (Figure 5). 18

The soil profile at the hillslope site was dry during summer with all the sensors measuring between 0 and 0.06 m<sup>3</sup>/m<sup>3</sup> of soil water content. In contrast the soil profile at the riparian site was relatively wet. At the riparian site, the soil water content at 0.1 and 0.3 m depth decreased to almost zero by March 2017 which indicates severe water stress for plants accessing water at these depths. However, the <u>riparian trees had access to water during the</u> <u>dry season at</u> relatively deeper soil from 0.5 to 0.9 m <u>depths\_depths, which had about 0.1</u> m<sup>3</sup>/m<sup>3</sup> of soil water<u>, which-This</u> was slightly higher than observed at the hillslope site at these **Commented [S6]:** Also address the query relating to ths statement either here or in the responses.

# 1 depths. This suggests that the riparian trees had access to some limited soil water at 0.5-0.9 m depths during the dry season.





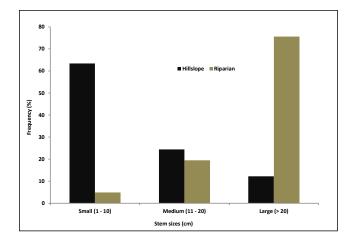
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Figure 5: Variations of soil water content at 0.1 to 0.9 meters depth on the hillslope andriparian sites.



# 6 3.2 Characterization of A. longifolia invasions

Tree heights on the hillslope ranged between 3 - 6 m, whereas they were 7 - 10 m at the 7 riparian site. The average trees counted in 100 m<sup>2</sup> quadrants were 18 and 25 on the hillslope 8 9 and riparian site respectively. Therefore, the average tree density was approximately 1 800 10 and 2 500 tree per hectare at the hillslope and riparian site, respectively. There was not much changes over time in canopy cover as A. longifolia is evergreen, and the leaf area index was 11 2.1 - 2.2 on the hillslope and 3.3 - 3.4 at the riparian site. The distribution of stem sizes at the 12 sites show that 63% of the stem sizes on the hillslope belonged to the small class size (1 - 10)13 14 cm) (Figure 6), whereas at the riparian site, most trees (70%) had large stem sizes (> 20 cm).



2 Figure 6: Distribution of stem sizes determined from measurements of stem diameters at

3 breast height of 40 trees on the hillslope and riparian site.

#### 4

#### 5 3.3 Transpiration dynamics

6 Transpiration rates of Acacia longifolia occurring at the riparian site were greater than those 7 on the hillslope throughout the June 2016 to June 2017 period (Figure 7). During the From August to December 2016-period, transpiration rates were generally high 2.0 - 3.5 mm/day at 8 the riparian site, and 1.0 - 1.4 mm/day at the hillslope site. From December 2016 to May 9 2017, transpiration rates generally decreased at both sites (Figure 7). The total water use by 10 A. longifolia during the period of the study on the hillslope was 242 mm/year and 596 11 mm/year at the riparian site. The riparian tree water use was 146 % greater than the hillslope 12 invasions. The 146% difference in water use rates was explained by high tree density 13 dominated by larger stem sizes in the riparian site. The differences in stem diameters and tree 14 density between riparian and hillslope sites have also been observed in other studies by Scott 15 16 (1999), Dye and Jarmain (2004), Clulow et al. (2011), and Dzikiti et al. (2013). Schachtschneider and Reinecke (2014) noted that plants growing under different conditions 17 18 of water availability can adapt their physiology to maximize their chances of survival. Therefore, the availability of water in the riparian zones caused trees to have large stem 19 20 diameters than those growing in non - riparian zones and thus increasing the detrimental effects of riparian trees to water resources. 21

2 The water use rates established in this study were similar to those determined in other studies

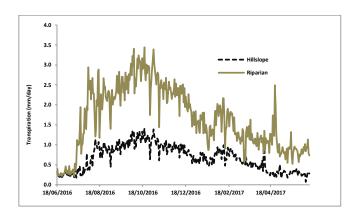
3 for Acacia species in the Western Cape. Dye and Jarmain (2004) estimated 171 mm/year, and

4 585 mm/year was estimated by Scott-Shaw et al. (2017). Other riparian invasions, such as

5 pines and eucalyptus used ~ 980 mm/year and ~ 833 mm/year respectively as reported by

6 Dzikiti *et al.*, (2013; 2016).





### 8

9 Figure 7: Seasonal variation of transpiration rates of *Acacia longifolia* at the hillslope and10 riparian sites.

11

Temporal variations of transpiration rates of *A. longifolia* invasions had similar patterns to the atmospheric evaporative demand (represented by  $ETo)_{a}$  which was the major driver of transpiration <u>during-from</u> winter to late spring when the trees accessed residual soil moisture from rainfall at both sites (R<sup>2</sup> = 0.6 and 0.5) (Figure 8a&b).

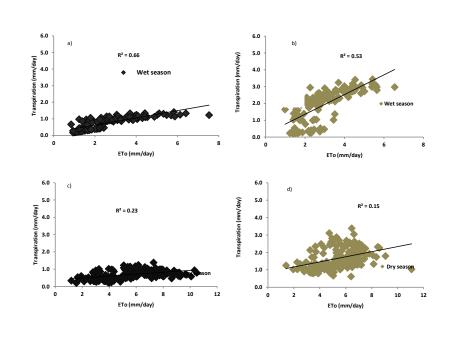


Figure 8: The relationship between transpiration rates of *Acacia longifolia* during the wet
season (June – September 2016) and dry season (October 2016 – March 2017) at the hillslope
(a & c) and riparian site (b & d).

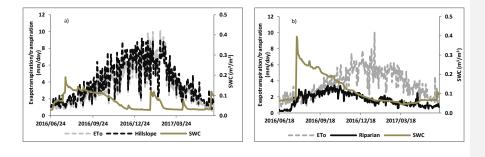
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5

1

6 During the dry periods soil water content declined from November 2016 to May 2017

7 causing transpiration rates of *Acacia longifolia* at both sites to decrease (Figure 9).



8

9 Figure 9: Seasonal variations of Acacia longifolia transpiration rates (Hillslope, Riparian),

10 reference evapotranspiration (ETo), and soil water content (SWC) at the study sites.

Figure 9 shows that while the reference evapotranspiration (ETo) rate was increasing from 1 October to December 2016, the transpiration rates of Acacia longifolia decreased as the soil 2 3 water content decreased. The decrease of the transpiration rate at the riparian site was rather 4 unexpected as it had been assumed that trees at this site would have access to groundwater occurring at depths greater than the monitored soil depth range (1.0 m). Five monitoring 5 boreholes developed in July 2017 for another related but separate study just outside the 6 7 Acacia longifolia stand at the riparian site had water tables at 2.4 to 2.9 depths below the surface. These findings therefore contradicted the previous literature (Scott, 1999; Doody et 8 9 al., 2011; and Nowell, 2011) which-that suggested that-riparian trees tend to strongly depend 10 on groundwater although different species from the one reported here were investigated in these studies. Snyder and Williams (2000) suggested that not all woody species in the forest 11 use groundwater and not all riparian trees benefit from seepage of water from the adjacent 12 river (Dawson & Ehleringer, 1991). The results of this study are in agreement with the 13 findings of Dzikiti et al. (2016) who concluded that water use by riparian Eucalyptus 14 15 declined with increasing soil water deficit in the upper soil horizons although the atmospheric evaporative demand was high. Thus rain water stored in the shallow soil layers was an 16 important source of water for the riparian eucalyptus trees which seem to be the case for the 17 Acacia longifolia monitored in this study. The A.longifolia trees at the riparian site had very 18 19 shallow root system (less than 60 cm) measured on trees uprooted by strong winds, which was evident from a number of trees that were uprooted by strong winds, due to poor anchorage. 20 The prevailing drought during the study period led to the substantial drying of the top soil 21 22 where most roots were concentrated.

23

### 24 3.4 Transpiration dynamics under unstressed conditions

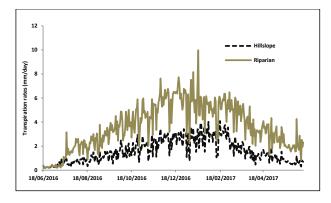
25 Rainfall data from the on-going monitoring in the catchment suggested that the total rainfall received in the catchment in 2017 was 50% of the long-term average for the catchment 26 27 (Mazvimavi et al., 2018). Therefore, total transpiration rates measured in 2016 and 2017 28 represented the water use rates when A. longifolia trees were water stressed. Unstressed transpiration coefficients (Allen et al., 1998) were estimated using data from August to 29 September 2016 when soil water content was not limited and the trees were transpiring 30 optimally. It was assumed that the unstressed coefficients remained the same from September 31 32 2016 to June 2017. The unstressed coefficient on the hillslope trees was estimated to be 0.39

#### 18

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- 1 for the hillslope and 0.90 for the riparian trees. These coefficients were used to estimate
- 2 unstressed transpiration rates (Figure 10).



5 Figure 10: Seasonal variations of transpiration rates of *A. longifolia* under unstressed6 conditions.

7

8 Under conditions where soil water is available for plants, riparian *A. longifolia* trees can use 9 up to 10 mm/day of water per day while the hillslope invasions can use up to 4 mm/day 10 (Figure 10). Under this scenario, water use on the hillslope can reach 579 mm/year and 1 348 11 mm/year at the riparian site. These results suggest that during years with rainfall above the 12 average, *A. longifolia* invasions can use more than two times the estimated rates in this study.

13 <u>These results are important to inform water resurces managers in catchments that</u>
14 <u>receive......(WRC report)</u>

15

# 16 3.5 The impacts of IAP's at a catchment scale

The effects on water resources was also estimated at catchment scale. This was based on the assumption that the area invaded by alien plants in the catchment is equivalent to 7 930 hectares at 100% density cover (Kotze *et al.*, 2010). It was also assumed that the riparian zone makes up 5% of the invaded area and 95% being the hillslope. If other alien species transpired at similar rates to *A. longifolia* trees, then the total water use by IAPs at a catchment scale would be 20.5 Mm<sup>3</sup> from June 2016 to June 2017<sub>a</sub> which was a dry year. The **Commented [Y10]:** Why are these findings important? Relate them to similar studies Acknowledge the limitations Make future recommendation

estimated water use by these plants is in close range with the estimated annual runoff (18.8 1 Mm<sup>3</sup>/yr) of the catchment (Mazvimavi et al., 2018). During years with rainfall above the 2 3 average, total water use by IAP's can reach 49 Mm<sup>3</sup>/yr. Transpiration rates of the natural 4 shrub lands were-was not investigated in this study. However, assuming that the fynbos shrubs transpired at similar rates to grasses, which use ~411 mm/year based on SAPWAT 4 5 estimates (van Heerden and Walker, 2016) - Then, then transpiration rates by IAP's were 30% 6 7 more than that of the fynbos shrubs. The hydrological benefits of clearing IAPs was estimated as the difference in transpiration rates between unstressed A. longifolia trees (579 8 9 mm/year) and the fynbos shrubs. Clearing IAPs and being replacedrestoring by with fynbos 10 plantations has the potential to result in 17 Mm<sup>3</sup>/yr of water being saved and available for other uses. These results are in agreement with the results by Nowell (2011) which that 11 suggested that the potential water savings after clearing IAP's at the Agulhas Plain over the 12 total area of 66 772 hectares was ~36 Mm3/yr. There will also be a gain in groundwater 13 recharge and/or stream flows if the alien trees are removed especially from the riparian zones 14 15 (Holmes et al., 2008; Scott-Shaw et al., 2017).

16 This study becomes one of the few studies that demonstrates that the water use by riparian

17 invasive species exceeds that by a similar species growing in non-riparian environment.

18 However, most studies, such as the current one, made measurements over small temporal and

19 spatial scales which makes it difficult to extrapolate findings (McConnachie et al., 2012).

20 Also, the successes and failures of alien clearing programmes have not been well

21 documented. This is why water resources managers have not consider alien clearing as the

22 first option to acquire "new" water.

# 23 4 Conclusions

24 Water use rates of A. longifolia growing in a flat riparian zone has been was found to be greater than when this species is growing on hillslopes. During the dry period, soil water 25 content was the main limiting factor for-to transpiration of A. longifolia at both the riparian 26 27 and hillslope sites. This contradicts the common assumption that riparian trees have ready 28 access to subsurface water. Generalizations about water use dynamics of riparian trees should therefore take into account the specific conditions or features of river water - riparian 29 subsurface water interactions, and rooting systems of the concerned species. Clearing of 30 invasive A. longifolia growing along the riparian zone should be prioritised due to the higher 31

water use rates in comparison to hillslope sites. Clearing of IAPs will contribute to making 1 "new" water available in the affected catchments. 2 3 Formatted: Font color: Text 1 Acknowledgement Formatted: Font: 12 pt 4 5 The study was funded by the South African Water Research Commission of South Africa, 4 Formatted: Font: (Default) Times New Roman, 12 pt 6 Formatted: Justified, Line spacing: 1.5 lines and the National Research Foundation/Applied Centre for Climate and Earth Systems 7 Formatted: Font: (Default) Times New Roman, 12 pt Science. Landowners and members of the Nuwejaars Special Management Area supported 8 9 this study by granting us permission to install monitoring equipment. 10 References 11 Allen, R.G., Pereira, L.S., Raes, D. and Smith, M., 1998. Crop evapotranspiration-Guidelines 12 for computing crop water requirements-FAO Irrigation and drainage paper 56. FAO, 13 Rome, 300(9), p.D05109. 14 Bailey, A. K. and Pitman, W. V., 2017. Water resources of South Africa. WRC Report No: 15 TT683/16/K5/2143, Water Research Commission, Pretoria. 16 17 18 Bullock, A., and Acreman, M., 2003. The role of wetlands in the hydrological cycle. Hydrology and Earth System Sciences, 7(3), pp. 358-389. 19 20 Bonanno, G., 2016. Alien species: to remove or not to remove? That is the 21 question. Environmental Science & Policy, 59, pp.67-73. Chamier, J., Schachtschneider, K. Le Maitre, D. C., Ashton, P. J. and Van Wilgen, B. W, 22 23 2012.. Impacts of invasive alien plants on water quality, with particular emphasis on South Africa. Water SA 38, pp. 345-356. 24 Clulow, A. D., Everson, C. S., and- Gush, M. B., 2011. The long term impact of Acacia 25 mearnsii trees on evaporation, streamflow, and groundwater resources. WRC Report no. TT 26 27 505/11, Water Research Commission, Pretoria.

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